

Research Article

Soil carbon stock in different land-use systems in the hilly terrain of Mizoram, Northeast India

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Abstract

Soil carbon is one of the most affected variables to land-use change in tropics. The soil carbon flux plays a major role in regulating microbial activities and nutrient distribution in soil. This study aimed to evaluate the soil carbon stock in various land uses at different depths in the hilly terrain of Mizoram, Northeast India. Soil samples at 0-10 cm, 10-20 cm and 20-30 cm soil depths were collected from Rubber plantation (RP), Oil palm plantation (OPP), Teak plantation (TP), Bamboo Forest (BF), 5 years fallow (5YF), 10 years fallow (10YF), *Tephrosia candida* plantation (TCP), Horticulture garden (HORT), Homegarden (HG) and Natural forest (NF). Soil carbon stock varied significantly ($p < 0.05$) across the land uses and depths. The soil under *Tephrosia candida* stand had significantly ($p < 0.05$) higher values of C stock (73.66 Mg ha^{-1}) which may be due to high biomass, dense vegetative cover and high C in root exudates. The minimum C stock estimated in Horticulture garden (43.28 Mg ha^{-1}) is probably due to reduced soil organic matter. Soil carbon stock in Homegarden, Teak plantation, Bamboo forest and Rubber plantation ranged from 46.82 Mg ha^{-1} to 59.34 Mg ha^{-1} whereas 5 years and 10 years fallow land, Natural forest and Oil palm plantation ranged from 61.35 Mg ha^{-1} to 73.35 Mg ha^{-1} . The study indicated that the land use change in the mountainous region significantly affected the carbon stock in the soil. A proper land use management strategies to increase the soil organic matter is recommended to enhance the carbon stock in this region.

Keywords: Land-use change, Mountainous region, Northeast India, Soil carbon stock, Soil depth

INTRODUCTION

Soil carbon is an important component of the global carbon cycle indicating soil fertility and productivity (Van der Werf *et al.*, 2009) and worldwide studies have shown a significant variation with relation to land uses (Ali *et al.*, 2017; Iqbal *et al.*, 2014; Maurya *et al.*, 2014). It is estimated that almost 60% of the world's terrestrial carbon is stored in forest vegetation and soil (McKinley *et al.*, 2011). 38% of the organic carbon is stored in the soil of tropical forests whereas 72% is found in soils of northern forests (Blais *et al.*, 2005). However, the growing population has drastically changed the land use pattern in tropical forests to meet their food and timber demands. Conversion of cropland to pasture land, tree plantation and secondary forest has shown a significant

increase in soil carbon stock (Guo and Gifford, 2002). Losses in soil carbon caused by the conversion of natural to cultivated vegetation are well documented (Manpoong and Tripathi, 2019; Yan *et al.*, 2012). Globally, 24% of the soil carbon is lost through forest conversion to cropland (Wei *et al.*, 2014).

Mizoram is a steeply sloped mountainous region with a geographical area of 21, 087 km² covering 86.27% forest and ranked second largest forest cover with respect to its geographical area in the country (Forest Survey Report, 2017). Land -use change is one of the major change processes and has altered the structure and functioning of forest ecosystems in the region. Tropical and subtropical forests of Mizoram are significantly shrinking due to change in land use pattern, particularly shifting cultivation and plantations which pro-

foundly affects soil properties (Manpoong and Tripathi, 2019; Oving et al., 2020). An increase in forest cover in certain areas is due to regeneration of bamboo species and other plantations (Forest Survey Report, 2017). A large extent of natural forest has turned up for Oil Palm and Rubber plantations as per land capability, suitability and sloppy terrain of the region (Economic survey Mizoram, 2016-2017).

Since, the land-use types affect the amount of litter input and the rate of decomposition (Lehmann et al., 2000), assessing soil as carbon sink in different land use systems, including forest ecosystem is important prior to land use management. Understanding the impact of land-use change on carbon stock is essential for effective land use planning and mitigation of climate change in the ecologically highly vulnerable hilly terrain of Mizoram. Thus, the present study aimed to assess the soil carbon stock in different land-use systems of hilly terrain of Mizoram.

MATERIALS AND METHODS

Study area

The studied land-use systems were situated in the northern part of Mizoram between 24°25'16" and 23°18'1" N Latitudes and 92°37'03" and 93°11'45" E Longitudes (Fig. 1). Mizoram constituting 0.64% of the country's geographical area comprises rugged, steep hill ranges and interspersed valleys. The state has different forest types belonging to 4 groups Tropical semi-evergreen, Tropical moist deciduous, Subtropical broadleaved hill and Subtropical Pine forests and is a part of Indo-Burma bio-geographic region, which is one of the richest biodiversity hotspots of the world. The state experiences the moist tropical to moist subtropical climate. During winter, the temperature varies from 11°C to 24°C and during summer it varies between 18°C to 29°C. The total annual rainfall varies from 2160 mm to 3500 mm. In terms of forest canopy density classes, the state has 131 sq km under very dense forest, 5861 km² under moderately dense forest and 12, 194 km² under open forest. The total carbon stock of forests in the state is 95.041 million tonnes (348.484 million tonnes of CO₂ equivalent) which is 1.34% of the total forest carbon of the country (Forest Survey Report, 2017).

Soil sampling

The soil samples were collected from 10 land uses: Rubber plantation (RP), Oil palm plantation (OPP), Teak plantation (TP), Bamboo Forest (BF), 5 years fallow (5YF), 10 years fallow (10YF), *Tephrosia candida* plantation (TCP), Horticulture garden (HORT), Homegarden (HG) and Natural forest (NF) to determine the variation in soil carbon (C) stock across different land use systems. Three sample plots of 20m × 20m

were randomly selected in each land-use system with five random sampling points in each sample plots. The vegetation, litter, roots, stones, and debris if any were removed from the surface of the sampling point. Soil samples were taken from three depths (0–10, 10–20 and 20–30 cm) to make a composite sample for each horizon. In this way, three composite replicates samples were taken from each land use. These samples were brought to the laboratory and then air-dried. The dried samples were lightly ground, sieved through a 2 mm mesh sieve and used to analyse soil organic carbon.

An undisturbed core sample was also collected from each plot at three depths (0–10, 10–20 and 20–30 cm) to estimate the bulk density by using a core ring and was packed in polyethylene bags and labeled.

Soil analysis

Soil C was analyzed by Walkley-Black chromic acid wet oxidation method (1934). Bulk density was determined by the core method described by Blake and Hartge (1986). The samples were oven-dried at 105°C for 24 hrs to constant weight and the oven-dry weight of the sample was divided by its volume in cubic centimeters to calculate the bulk density of the soil. The soil samples were then dry sieved and weighed to obtain <2 mm size fractions and >2 mm rock fragments. The soil C stock (Mg ha⁻¹) was calculated by multiplying the soil C concentration (%) with the calculated bulk density (g cm⁻¹) and depth of soil layer (cm) by using the following formula described by Batjes (1996).

$$\text{Soil C stock (Mg ha}^{-1}\text{)} = \text{soil C (\%)} \times \text{BD (g cm}^{-1}\text{)} \times \text{soil depth (cm)} \times (1 - F) \quad \dots \text{Eq. 1}$$

where,

F is the volume of the fraction of fragments > 2 mm

BD-Bulk density

Statistical analysis

An analysis of variance was carried out to determine the effect of land use on bulk density, soil C, and soil C stock at different soil depths. A least significant difference (LSD) comparison test was done to separate statistically different means ($p < 0.05$). SPSS version 18.0 (SPSS Inc., Chicago, IL) was used for the statistical analysis.

RESULTS AND DISCUSSION

Effect of land-use change on soil carbon

Soil C differed significantly ($p < 0.05$) between land-use types at all the studied soil depths (Table 1). Soil C concentration was maximum in NF which was followed in decreasing order by TCP > 5 YF > OPP > 10 YF > RP > BF > HG > HORT > TP. The magnitude and direction of soil C dynamics are gradually affected by multiple factors, including climate, soil type, soil depth,

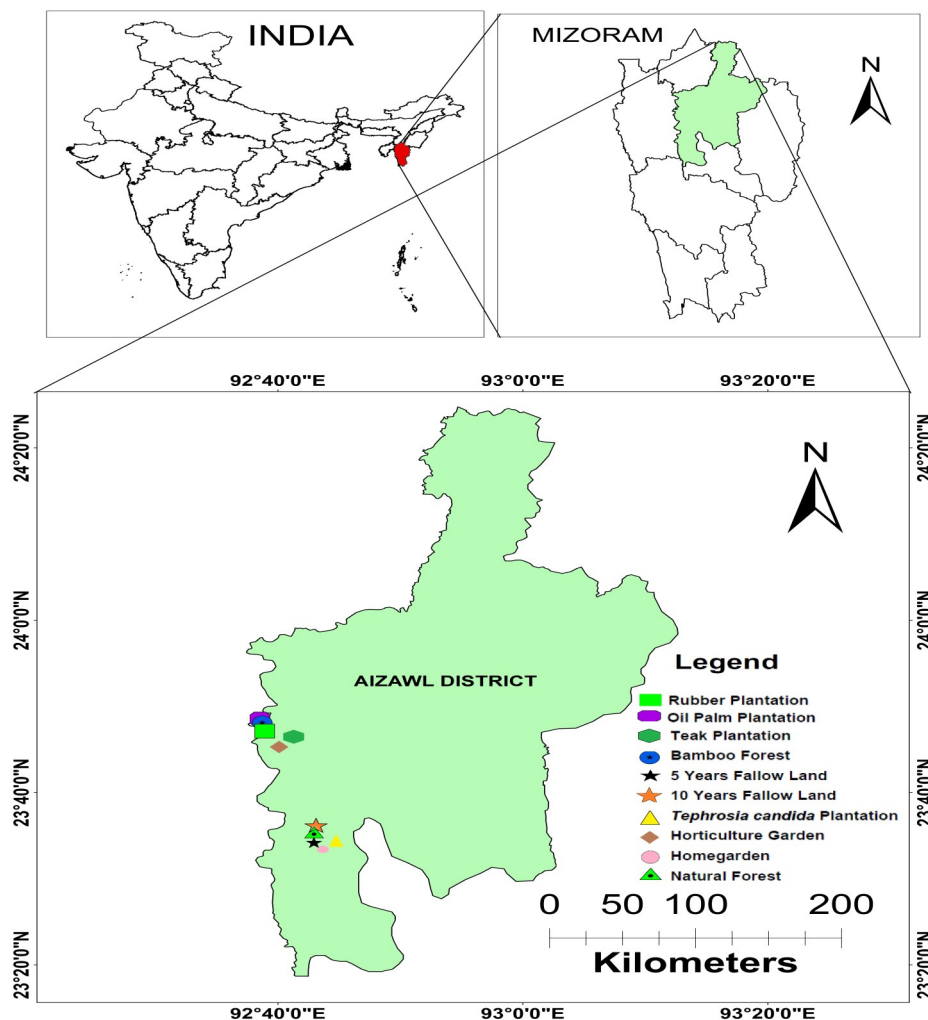


Fig. 1. Map of study area of Aizawl District of Mizoram.

tree species, nutrient management and vegetation (Don *et al.*, 2011; Wapongnungsang *et al.*, 2018). Higher soil C in NF could be due to the greater biomass input through vegetation as soil organic matter generally has a positive relationship with plant density. The reduced inputs of C into the soil compared to the native land use subsequently increases the rate of soil organic matter decomposition, which possibly led to a net loss of soil C in other land uses. Similar findings have been reported during the conversion of farmland to the plantation, secondary forests and grasslands of China by Zhang *et al.* (2010).

Effect of land-use change on soil C stock

The soil C stock was found maximum in TCP which was followed by OPP > 10 YF > NF > 5 YF > RP > BF > TP > HG > HORT (Table 1). Conversion of NF led to sharp decrease of C stock in 5YF, RP, BF, TP, HG and HORT (8.8%, 11.8%, 29.4%, 30.3%, 30.4% and 35.7% respectively). Comparatively, there was a small increase in soil C stock in 10 YF, OPP and TCP (3%, 8.9% and 9.4 %, respectively) with the change in land

use. Guo and Gifford, (2002) also reported a decline in soil C stocks after land use changes from pasture to plantation (-10%), native forest to plantation (-13%), native forest to crop (-42%), and pasture to crop (-59%). Significant differences of soil C stock between land uses could be due to the variation in quality and quantity of aboveground biomass (which led to different C concentrations) as well as variation in soil bulk density under different land use systems. The maximum soil C stock in TCP is possibly due to high biomass, dense vegetative cover, high C in root exudates and a deep root system with high nitrogen fixing ability (Manpoong *et al.*, 2020a). *Tephrosia candida* plantation has played a vital role in restoring degraded land and shifting cultivation sites to enhance soil fertility and crop productivity (Wapongnungsang *et al.*, 2017). A distinct increase in soil C stock was recorded in 10 years fallow compared to 5 years of fallow land. This could be attributed to the progressive accumulation of soil organic matter and the recovery of soil microbial communities with the increase in the fallow period (Wapongnungsang *et al.*, 2018; Manpoong *et al.*, 2020b). A longer fallow period

Table 1. Soil C stock (Mg ha⁻¹) in different land-use systems. LSD is shown at $p < 0.05$.

Land-use system	Depths	Soil carbon %	Bulk density (g cm ⁻³)	Soil C stock (Mg ha ⁻¹)
Rubber Plantation	0-10 cm	1.62	1.33	21.43
	10-20 cm	1.43	1.40	19.94
	20-30 cm	1.23	1.47	17.97
	0-30 cm	1.42	1.40	59.34
	LSD	0.20	0.04	2.93
Oil Palm Plantation	0-10 cm	2.15	1.25	26.96
	10-20 cm	1.91	1.36	25.99
	20-30 cm	1.46	1.40	20.40
	0-30 cm	1.84	1.34	73.35
	LSD	0.17	0.08	3.03
Teak Plantation	0-10 cm	1.08	1.30	14.00
	10-20 cm	1.04	1.53	15.87
	20-30 cm	1.01	1.69	17.00
	0-30 cm	1.04	1.50	46.87
	LSD	0.34	0.17	5.6
Bamboo Forest	0-10 cm	1.53	1.08	16.51
	10-20 cm	1.33	1.26	16.64
	20-30 cm	1.03	1.39	14.32
	0-30 cm	1.30	1.24	47.47
	LSD	0.33	0.03	4.36
5 years Fallow	0-10 cm	2.26	0.90	20.37
	10-20 cm	2.16	1.04	22.47
	20-30 cm	1.63	1.14	18.51
	0-30 cm	2.01	1.03	61.35
	LSD	0.16	0.14	1.99
10 years Fallow	0-10 cm	1.99	1.04	20.65
	10-20 cm	1.80	1.36	24.57
	20-30 cm	1.67	1.44	24.13
	0-30 cm	1.82	1.28	69.36
	LSD	0.23	0.20	5.85
<i>Tephrosia candida</i> Plantation	0-10 cm	2.37	1.09	25.68
	10-20 cm	2.24	1.16	25.99
	20-30 cm	1.80	1.22	21.99
	0-30 cm	2.14	1.16	73.66
	LSD	0.22	0.15	7.05
Horticulture garden	0-10 cm	1.19	1.17	13.97
	10-20 cm	1.18	1.24	14.34
	20-30 cm	1.08	1.38	14.97
	0-30 cm	1.15	1.27	43.28
	LSD	0.19	0.42	4.17
Homegarden	0-10 cm	1.43	1.24	17.72
	10-20 cm	1.10	1.38	15.16
	20-30 cm	0.97	1.43	13.94
	0-30 cm	1.17	1.35	46.82
	LSD	0.17	0.20	3.39
Natural Forest	0-10 cm	2.42	0.92	22.25
	10-20 cm	2.21	0.99	21.88
	20-30 cm	1.95	1.19	23.18
	0-30 cm	2.19	1.03	67.31
	LSD	0.13	0.08	1.40

leads to the recovery of original soil status, with a gradual increase in soil C and improved microbial structure (Aboim *et al.*, 2008). Similarly, Laganieri *et al.* (2010) reported that with an increase in time, there is an increase in the quantity of carbon inputs, accompanied by a new microclimatic regime and enhanced organic matter that promotes C accumulation in the soil. The

lowest soil C stock in TP, HG and HORT could be due to low organic matter inputs through the existing vegetation. Several workers have reported that natural forests store more carbon than any other terrestrial ecosystem, thereby representing a significant carbon pool for the global carbon budget (Manpoong and Tripathi, 2019; Yang *et al.*, 2018; Haghdoost *et al.*, 2013). Re-

cently, Singh *et al.* (2018) reported the highest soil C stock in agroforestry land use (50.85 Mg C ha⁻¹) and the lowest in agricultural cropland (33.99 Mg C ha⁻¹) when compared with forest and plantations of Mizoram.

Effect of soil depth and land-use on soil C and soil C stock

Soil C decreased with depth and bulk density increases with depth, whereas varied trends were observed in soil C stock (Table 1). The highest soil C at upper depth can be attributed to the continuous addition of organic debris, partially decomposed plant and animal remains and micro faunal population in the top soils. Several workers have reported that the upper soil horizon accumulates the maximum organic materials than the deeper horizon (Wang *et al.*, 2008; Emiru and Gebrekidan, 2013).

Soil C stock in RP, OPP and HG decreased with increasing depth whereas in BF, 5 YF, 10 YF, TCP the maximum soil C stock was found at 10-20 cm depth. In the present study soil C stock at different depths was found to be very sensitive to land use types. Maximum soil C stock in RP, OPP and HG at upper depth suggested that the surface soil is more active for sequestering C from the atmosphere after the land use change. Haghdoost *et al.* (2013) reported the same trend during conversion of natural forest to other land-uses in northern Iran and interpreted to occur as a result of slow litter decomposition and humus conversion that starts from the soil surface.

The maximum soil C stock at 10-20 cm depth in jhum fallow lands is possibly due to crop diversification's long-term effect during shifting cultivation that causes differences in quality of aboveground litter inputs and decomposition. The alteration in decomposition dynamics due to diversifying crops in agricultural systems has been reported by McDaniel *et al.* (2016). Arevalo *et al.* (2009) argued that about 60% of soil C in tropical soils is stored within 0-15 cm soil depth. However, TP, HORT and NF recorded the maximum soil C stock at 20-30 cm soil depth, possibly due to higher bulk densities. The study on forest plantation in eastern Australia showed that carbon accumulation and loss patterns also vary according to location, tree species and plantation management system (Turner *et al.*, 2005). In addition, changes in temperature and precipitation pattern may also have a major influence on the decomposition and amount of soil C stored within tropical ecosystems.

Conclusion

Land use and soil depth had a significant ($p < 0.05$) impact on bulk density, soil C concentration and soil C stock. Soil C was negatively correlated with bulk densi-

ty across the land-use systems. The soil C stock varied both in terms of soil depth and different land use. The maximum soil C stock in *T. candida* plantation was probably due to high root biomass and C exudation rate, whereas Horticulture garden reported the minimum among the studied land use. The greatest fluxes of soil C resulted from the conversion of native forests to other land use. Therefore, maintaining adequate soil organic matter content, proper soil structure and other soil properties are obligatory for the sustainability of the land-use systems. The present study suggests that *T. candida* can play a vital role in improving the soil C stock in degraded land uses of Mizoram. A further study on the rate of soil organic matter decomposition across the land-uses will improve the understanding of soil carbon dynamics along with a soil profile and the potential response of soil carbon to land-use change.

Conflict of interest

The authors declare that they have no conflict of interest.

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